Application of a 3D Lagrangian model to explain the decline of a *Dinophysis* 1 2 acuminata bloom in the Bay of Biscay 3 Velo-Suárez, L.^{1*}, Reguera, B.¹, González-Gil, S.², Lunven, M.³, Lazure, P.³, Nézan, E.⁴, 4 5 and Gentien, P.³ 6 ¹ Instituto Español de Oceanografía, Centro Oceanográfico de Vigo, Aptdo. 1552, E-7 8 36200 Vigo, Spain ² Instituto Español de Oceanografía, Centro Oceanográfico de Madrid, Orense 58, 7^a 9 10 planta, E- 28020 Madrid, Spain ³ IFREMER, Centre de Brest, DYNECO, Pointe du Diable BP70, E-29280 Plouzane, 11 12 France ⁴ IFREMER, 13, rue de Kérose, Le Roudouic E-29187 Concarneau Cedex, France 13 14 15 * Corresponding author: 16 Lourdes Velo-Suárez 17 Instituto Español de Oceanografía 18 Centro Oceanográfico de Vigo Aptdo. 1552, E-36280 Vigo, Spain 19 20 Tel: 34-986-492111 21 Fax: 34-986-498626 22 e-mail: lourdesvelo@gmail.es 23 24

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During July 2006, a cruise was carried out in the Northern Bay of Biscay (off Brittany, France) to study meso-and microscale patterns of phytoplankton distribution. Special attention was focused on the fine scale vertical distribution of *Dinophysis* spp. and its physiological status. Moderate concentrations $(10^2-10^3 \text{ cells L}^{-1})$ of *Dinophysis* acuminata were constrained to specific depths (upper layers of the pycnocline) at stations with lower surface salinity (34.5) and steep temperature gradients (18-13.5°C between 04 and 07 m depth) within the Loire and Vilaine river plumes. On board observations showed a D. acuminata population at the maxima with 89% of viable (FDA-treated) cells and moderate growth (up to 0.23 d⁻¹) in cell isolates that showed a positive growth response to DOM (M.W.>1 kD) from the same area concentrated by ultrafiltration. Despite the good physiological conditions of the cells, the population of *D. acuminata* declined rapidly to undetectable levels during the second leg of the cruise. For the first time a 3D Lagrangian Particle Tracking Model was used to explore the effect of physical dispersion in the decline of a *Dinophysis* bloom in the Loire and Vilaine river plumes (Bay of Biscay) during the exceptionally hot summer of 2006.

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45 *Keywords:* Lagrangian individual particle tracking models, harmful algal blooms,

46 *Dinophysis acuminata*, physical-biological interactions, Bay of Biscay.

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1. Introduction

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50 Diarrhetic shellfish poisoning (DSP) is a gastrointestinal disease resulting from 51 ingestion of shellfish contaminated with OA-related lipophilic dinoflagellate toxins. 52 These toxins threaten public health and shellfish industries due to its worldwide 53 occurrence, high morbidity rate and the long duration of shellfish toxicity outbreaks even at very minute (< 10² cells L⁻¹) cell concentrations (Yasumoto et al., 1985). 54 55 The French Phytoplankton Monitoring Network (REPHY) was created by the 56 IFREMER (Institut Français de Recherche pour l'Exploitation de la Mer) in 1984 to 57 detect the seasonal occurrence of potentially toxic dinoflagellates, with emphasis on 58 Dinophysis species. Since then, seasonal blooms of D. acuminata associated with DSP 59 outbreaks have been nearly annual features in different regions of the Bay of Biscay 60 (France) (Delmas et al., 1992). 61 Unveiling the nutritional sources for *Dinophysis* spp. has been a challenge for decades. 62 Nevertheless, recent work has shown D. acuminata to be a mixotroph that grows well 63 when feeding on the photosynthetic ciliate *Mesodinium rubrum* (= *Myrionecta rubra*) 64 (Park et al., 2006). Recent culture results have suggested that D. acuminata is an obligate 65 mixotroph that requires light, nutrients and live ciliate prey for long-term survival (Kim 66 et al., 2008; Riisgaard and Hansen, 2009). Evidences of D. acuminata blooms promoted 67 by elevated inorganic nutrient concentrations have not been found (Delmas et al., 1992). In contrast, a co-occurrence of *Dinophysis* maxima and that of organic matter aggregates 68 69 has been reported by several authors (Gentien et al., 1995; Lunven et al., 2003).

Different mechanisms have been suggested for the control of the initiation and development of dinoflagellate blooms. Nevertheless, little is known about the factors that trigger the decline of *Dinophysis* blooms. Turbulence and currents may act to dissipate or concentrate the D. acuminata cells while the population numbers may change due to intrinsic features (division, mortality) and physical-biological interactions (see Velo-Suárez et al., 2009). In the coastal waters of the Northern Bay of Biscay, D. acuminata has been found to exhibit a very heterogeneous vertical distribution and to concentrate around water density gradients (Gentien et al., 1995; Maestrini, 1998). Different studies in the region (Lunven et al., 2005; Marcaillou et al., 2001; Lunven et al., 2005) have shown horizontal confinement of *D. acuminata* blooms in river plumes. Delmas et al. (1993) suggested a transport of D. acuminata populations by water inflow from offshore to coastal areas of the Northern Bay of Biscay. Following this hypothesis, Xie et al. (2007) used a 3D model under realistic forcing conditions to test the role of retention areas and their subsequent advection to the coast in promoting DSP coastal outbreaks caused by *Dinophysis* populations. 3D Lagrangian particle-tracking models (3D LPTM) and individual based models (IBM) have been used by the oceanographic community during the past two decades to explore processes that influence the transport of eggs and developing larval stages of invertebrates and fish(Gallego et al., 2007). Results from these studies have shown how this kind of models has been able to explain the transport/retention of fish larvae in some regions (Santos et al., 2007. 3D-LPTM models can provide new insights into physicalbiological interactions that affect plankton dispersal, growth and survival and enhance our understanding of plankton population variability and structure (Gallego et al., 2007).

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During summer 2006, a three-week multidisciplinary cruise was conducted to describe the fine-scale physical and biological structure of the water column and to study meso-and microscale patterns of phytoplankton distribution in the Northern Bay of Biscay. For this purpose, we used outputs from a realistic hydrodynamic model to force a 3D LPTM, and these simulation results are discussed together with the *D. acuminata* observations made during the survey. This work represents the first application of a 3D LPTM model to study the role of physical dispersion in the decline of a harmful dinoflagellate (*D. acuminata*) bloom located in the Loire and Vilaine river plumes, off Brittany (Bay of Biscay).

2. Material and methods

104 2.1. The study area

The Northern Bay of Biscay is an open bay off the West coast of France (Fig 1).

Freshwater discharges from the Loire and the Vilaine Rivers into the Bay create a long-shore flow that may dominate a wide part of the continental shelf during the high-runoff season (from November to April). Propagation of the Loire and Vilaine river plumes are responsible for a marked haline stratification over the shelf that progressively weakens in

summer—with the decrease in river run-off—and is the weakest at the beginning of

autumn (Lazure and Jegou, 1998).

Apart from this highly variable and seasonal haline stratification, weak tidal currents in the region allow thermal stratification to be established over the entire continental shelf from spring to early autumn. Winds in the Brittany coast show an annual cycle whose

115 main characteristic is the spring (late March) shift from south-westerly to west-116 northwesterly winds (Lazure et al., 2008). 117 118 2.2 REPHY data 119 Biweekly concentrations of D. acuminata throughout the year were obtained from 120 REPHY (IFREMER) database (http://www.ifremer.fr/envlit/surveillance/rephy.htm). 121 Phytoplankton quantitative analyses were carried out according to the Utermöhl (1931) 122 method. Samples were counted after sedimentation (4 h minimum) of 10 mL columns 123 under an inverted light microscope (IX70 Olympus) fitted with 10X, 20X and 40X 124 objectives and phase-contrast optics. 125 126 2.3 Cruise sampling overview 127 Studies were carried out on board R/V *Thalassa* from 06 to 22 July 2006. Sampling 128 stations from the first (06-12 July) and second (13-22 July) leg of the cruise are shown in 129 Fig 2 and Table 1. At each station, vertical plankton net (20-µm mesh; 20 m depth) hauls 130 were collected and immediately examined on board under a Nikon ECLIPSE 2000 TE-S 131 inverted microscope at 100X and 400 X magnification. 132 Measurements in the water column were carried out with the high-resolution 133 IFREMER particle size analyzer profiler (IPSAP), capable of simultaneous resolution of 134 physical and optical structures at small scale (see Gentien et al., 1995; Velo-Suárez et al., 135 2008 for details). The IPSAP profiler includes a fluorescence sensor (Seapoint Sensors, 136 Inc., Exeter, New Hampshire, USA) attached to a SBE25 CTD probe (Sea-Bird 137 Electronics, Washington, USA) coupled to an in situ particle-size analyzer (Gentien et al., 1995). The latter is needed as guidance for sampling, since it detects specific particle profiler, helping to distinguish between layers with abundant organic aggregates and those with dinoflagellates in the 36-64 µm fraction. The SBE25 probe allows real time data acquisition of standard parameters, such as depth, temperature, salinity, chlorophylllike in vivo fluorescence and photosynthetically active radiation (PAR). The CTD fluorometer was calibrated with laboratory cultures of the diatom *Chaetoceros gracile* using the trichromatic method for chlorophyll determination according to Aminot and Kerouel (2004). IPSAP data were post-processed to obtain averages every 50-cm. Data presented thereafter are 50-cm averages without any overlapping. A rosette (12 5-L Niskin bottles) attached to the IPSAP profiler was used to sample at the exact depth of the sensors. Two Niskin bottles were casted and closed at each desired depth during the upcast. After retrieval of the profiler, water samples from each bottle were immediately well mixed and then subsampled for biological and chemical analysis. From each depth, a 2-L sample was gently concentrated through a 20-µm mesh and filtered material resuspended in a volume of approximately 20 mL; 3 mL of this concentrate were poured into a sedimentation slide for in vivo observations and for a quick rough estimate of *Dinophysis* abundance. For quantitative analyses of microphytoplankton, two kinds of samples were taken: (1) unconcentrated and (2) 2-L seawater samples concentrated through 20-µm filters, that were further resuspended in a final volume of 30 mL (factor = 66.6). Samples were preserved in acidic lugol and analyzed under an inverted microscope (Nikon ECLIPSE 2000 TE-S) according to the Utermöhl (1931) method. Phytoplankton abundance was determined to species level when possible. Concentrations of smaller and more abundant species were estimated

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from two transects counted at 400X magnification (detection limit: 40 cell L^{-1}). In the case of less abundant species (including *Dinophysis* spp.), specimens from the whole chamber were enumerated at a magnification of 100X after sedimentation of 3-mL aliquots of the concentrated samples. In the last case, the detection limit in cell counts was 5 cells L^{-1} .

2.4 D. acuminata in vivo observations and viability

On board observations of the physiological status of *D. acuminata* from live samples were performed under a Nikon ECLIPSE 2000 TE-S inverted microscope (Nikon Instruments Europe B.V., The Netherlands) with Differential Interference Contrast (DIC) and epifluorescence under a blue filter set (excitation 450-490 nm, emission 520 nm LP). Snapshots were taken with a Nikon D70 camera coupled to the microscope at magnifications of 100X and 400X. Observations included annotations on the chlorophyll (red) and phaoephytin (orange) autofluorescences in *D. acuminata* and detection and frequency of cells parasitized by *Amoebophrya* spp., which exhibited a characteristic green autofluorescence

Viability assays of *D. acuminata* cells were carried out with specimens stained with

fluorescein diacetate (FDA; Sigma Chemicals, St Louis, MO, USA). Previous studies with different vital/mortal stains (i.e SYTOX-Green, Trypan Blue, Calcein-AM) had shown FDA to give the best results with natural populations of *D. acuminata* (Gónzalez-Gil, Personal communication). The non-polar and non-fluorescent FDA molecules enter freely into the cells and act as substrate for non-specific cell esterases. This reaction results in the formation of polar and fluorescent molecules that can be retained within the

184 intact cell membrane of live cells (Coleman and Vestal, 1987). In this way, FDA stains 185 all cells with active esterases and intact cell membranes, i.e. viable cells. The FDA stock solution of 5 mg mL⁻¹ was dissolved in dimethyl sulfoxide (DMSO) and 186 187 stored at -20°C. A portion of the previously concentrated live sample (ca. 3 mL) was 188 transferred to a sedimentation chamber and stained with 5-10 µL of the FDA working 189 solution (5 µM final concentration) and incubated in the dark for 15 min. Viable cells 190 showed green fluorescence after excitation with blue light. Red fluorescence was not 191 considered as an indication of viability since chlorophyll a in dead cells may 192 autofluoresce after cell death (Gentien, 1986). Percentages of positive and negative 193 binding were estimated after scanning of 100 cells at least of *D. acuminata* whenever 194 possible. 195 196 2.5 Incubations with Dissolved Organic Matter (DOM) enrichment. 197 A high flow-rate peristaltic pump on board connected by a hose to the profiler was used 198 to pump large volumes (1000 L) of seawater from the depth of the particle maximum 199 depth (18.9 m) at station 9 (see Fig 2 and Table 1 for location and date) for the DOM 200 extraction by ultrafiltration). Water was pre-filtered through 3-and 1-µm Millipore filter 201 cartridges, and then onto GF/C, GF/F (Whatmann), 0.45 µm nucleopore filters and 0.2 202 μm Opticap filter units (Millipore). DOM was concentrated by means of the Prep/scaleTM 203 TFF 6ft2 cartridge (Millipore) tangential ultralfiltration device (Benner et al., 1997; Guo 204 et al., 1994; Purina et al., 2004). 205 For the incubation experiments, 5 L of seawater were concentrated by reverse filtration 206 (Dodson and Thomas, 1978) through a 20-µm mesh to a final volume of about 250 mL

and maintained in the incubation chamber on board—at 15±1°C with a 16:8 light:dark photoperiod—in glass cylinders coated with aluminium foil, except the last few centimetres near the top, to enhance accumulation of swimming dinoflagellates in the upper illuminated portion. In that way, actively swimming (healthy) cells of D. acuminata concentrated in a cloud near the surface as described by Maestrini et al. (1995). Dinophysis acuminata cells, concentrated in the upper portion of the cylinders, were siphoned out and individually isolated with microcapillary pipettes under 25X and 100 X microscope magnification. Each cell was transferred two to three times to slides with filtered-sterilized seawater (0.2 µm Millipore filters, Millipore Corp. Bedford, MA) and finally placed (individual cells and groups of 5 cells) in culture-plate wells of 0.2 mL containing 150 µL of either filtered seawater (FSW) from the same depth of collection (control) or the same water enriched with dissolved organic matter of M.W.>1 kD (DOM>1 kD) to a final concentration 50 times higher than the background level. Incubations were examined daily to record the number of cells on each well. Estimates of division rates (µ) were obtained according to:

$$\mu = \frac{\ln N_f - \ln N_0}{t} \qquad (1)$$

Where μ (d⁻¹) is the specific division rate, N_f and N_0 are the final and initial cell numbers (cell/well) and t the time (d).

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2.6 The model

The MARS3D hydrodynamical model (Lazure and Dumas, 2008) for the Northern

Bay of Biscay was used to simulate the hydrodynamics in the study area. Model

simulations extended from the French coast to 8°W and from the Northern Spanish coast

230 to Southern England. The horizontal mesh size used had a resolution of 4 km, and 30 231 vertical levels were considered with a finer resolution near the surface. The simulation 232 was forced by real conditions of tide, river run-off, heat fluxes and wind using data from 233 the Arpège model of the French Met Office. Spin up from one year (2005) was 234 considered. 235 A Lagrangian individual-particle tracking 3D-model—Ichthyop (Lett et al., 2008)— 236 was used to simulate the possible advection or dispersion of virtual *Dinophysis* 237 populations in the study area. Environmental conditions (3D field of currents, 238 temperature and salinity) between 15 June and 30 July were provided by archived 239 simulations of MARS3D configured for the Northern Bay of Biscay. In the Bay of 240 Vilaine (BV), two areas were defined, Area 1, which extended from 47.8° N to 47.2° N 241 and from -2.8° E to -2.1° E and roughly corresponds to station 31 from the first leg of the 242 cruise (Fig 2) and Area 2, which represented the whole BV (Fig 1), from 46.7°N to 47.8°N and from 2° W to 3.3° W. Virtual particles movement was defined as a passive 243 244 Lagrangian transport. 245 In this work, two sets of simulations were carried out. In the first set, each simulation 246 consisted of a random release of 5000 particles representing virtual cells of *Dinophysis* 247 within Area 1 every 5 days from 15 June to 30 July (0-20 m). Each simulation lasted 10 248 days. The proportion of particles retained was estimated every 2 m depth from surface (0) 249 m) to 20 m. Particles were considered to be retained at the end of each simulation if they 250 were still present in Area 1 or Area 2. 251 In the second set, Ichthyop simulations were used to evaluate the trajectory and the 252 retention-dispersion pattern of 1000 virtual particles, from 9 to 19 July, at 2-5 m depth, i.e.

the depth range where D. acuminata cells were detected during the first leg of the survey 253 254 (see section 3.5). 255 256 3. Results 257 3.1 Seasonal distribution of Dinophysis during 2006 258 The seasonal occurrence of *Dinophysis* spp. in the BV in 2006 showed a rather wide 259 temporal distribution. The *Dinophysis* growth season as defined in Xie et al. (2007) 260 started by the end of April (Fig 3) and two weeks later, on 15 May 2006, Dinophysis populations reached 2300 cell L⁻¹. The bloom lasted until the end of July and the highest 261 cell density (6200 cell L⁻¹) was recorded on 19 June 2006. Therefore, the cruise (05-22) 262 263 July) was carried out two weeks after the seasonal maximum and when *Dinophysis* 264 numbers were decreasing. 265 266 3.2 Dinophysis spp. distribution during the cruise 267 During the first leg of the cruise (06-12 July), concentrations of *Dinophysis* spp. were extremely low ($< 50 \text{ cells L}^{-1}$) offshore but higher values (up to 500 cells L⁻¹) were found 268 269 in the River Loire and Vilaine plumes at stations with a high particle load, lower surface 270 salinity and steep vertical gradients of temperature (Fig 4, Fig 5). Sea surface temperature 271 (SST) estimated from satellite images presented a positive anomaly—in relation to a 22-y 272 (1986-2008) time series; courtesy of Nausicaa, IFREMER)—of 2.98°C on 12 July (Fig 273 6). Temperature values obtained from the CTD casts exceeded 19°C at surface and 274 exhibited strong horizontal and vertical gradients along the study area (Fig 5B, 5E).

275	Vertical distribution of <i>D. acuminata</i> in BV showed that the population was located
276	between the surface and the beginning of the pycnocline (1-4 m depth; Table 2, Fig 5)
277	within a community dominated by other dinoflagellates (Prorocentrum micans, Ceratium
278	fusus and Gonyaulax diacantha). D. acuminata maxima were never found associated to
279	the chlorophyll a (Chl a) maxima. The maximum concentration (1002 cells L^{-1}) was
280	observed on station 31 at 2.25 m depth (T= 18.5°C, S= 33.8 σ_t , F equiv. to 0.85 μg Chl a
281	L^{-1}) on 11 July (Fig 7). At this point, the phytoplankton community located at the Chl a
282	maximum (15 m; F=14.5 μ g Chl a L ⁻¹) was composed by a population of
283	Chrysochromulina sp. $(1.5 \times 10^6 \text{ cells L}^{-1})$ —that developed into a dense thin layer during
284	the second leg of the cruise—a diatom assemblage that comprised a mixture of
285	Leptocylindrus danicus, L. minimus and medium-sized Chaetoceros spp. ($< 2x10^5$ cell
286	L^{-1}) and a dinoflagellate assemblage dominated by <i>Prorocentrum</i> spp. and ecdysal cysts
287	of Fragilidium spp. (ca. 5000 cell L^{-1}).
288	The horizontal distribution of <i>D. acuminata</i> during the second leg of the cruise (13-22
289	July) (Fig 8) shows that unlike during the first leg, concentrations of D. acuminata during
290	this period were very low ($< 100 \text{ cells L}^{-1}$) in the whole study area.
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292	3.3 On board observations of D. acuminata
293	No parasites were seen in any of the live samples observed by epifluorescence.
294	Most D. acuminata cells in FDA-treated samples appeared to be in good condition, with
295	branched, brightly autofluorescent (orange) chloroplasts. The frequency of viable cells
296	was higher than 89% in all samples (8) analyzed. Therefore, a predominance of green-

297	fluorescing metabolically-active cells was observed although some non metabolically-
298	active cells with weak red v or no fluorescence at all were also found (Fig 9).
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300	3.4 Short-term incubations with DOM enrichments.
301	Individually picked cells of D. acuminata incubated in the well-plates showed a
302	positive growth, both in the filtered seawater controls and in the wells enriched with
303	concentrated DOM. Until day 4, D. acuminata cell-numbers increased in the controls and
304	in the DOM-enriched both experimental treatments in a similar manner, but from day 5
305	onwards, cells in the control wells stopped dividing, whereas those in the DOM-enriched
306	treatment continued division until the end of the experiment (day 7; Fig 10).
307	However, in comparison with the controls (max. 0.10 d ⁻¹), cells incubated in DOM-
308	enriched water exhibited a significantly higher division rate (up to 0.23 d ⁻¹) (Fig 10).
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310	3.4 Other biological observations
311	An interesting observation during the first leg of the cruise was the frequent detection of
312	Fragilidium spp., at times the dominant or co-dominant microplanktonic species at the
313	depth of the chlorophyll- or the particle maximum. Probably due to the stress of
314	concentration procedures, cells appeared mainly as ecdysal cysts. After dissection of
315	individual specimens, a new species of Fragilidium—F. duplocampanaeforme—was
316	described and observed to contain preyed cells of D. acuminata (Nézan and Chomérat,
317	2009).
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319	3.5 Modelling simulations

320	Results from the Ichthyop model simulations showed that, on average, particle-retention
321	on Area 1 increased with depth, from 0-10% at a releasing depth of 0-10 m to 40-50% at
322	20 m (Fig 11A). The deeper the particles were released, the more retention over the Area
323	1 was observed (Fig 11A). The estimated retention in this Area was higher from 25 June
324	to 5 July and from 15 to 30 July. In contrast, minimum values of retention were observed
325	at 0-5 m depth on 20 June and 10 July. These simulations showed that less than 10% of
326	the particles would be retained in Area 1 after 5 days, between 20 and 25 June.
327	Field observations showed that maximum concentrations of <i>D. acuminata</i> were found in
328	Area 1 on 19 June and 11 July. On average, retention of particles released in the upper
329	layer 0-5 m—the depth range where <i>D. acuminata</i> populations were detected—in Area 1
330	was maximal on 25 June-05 July and minimal on 20-25 June and 10-15 July (Fig 11).
331	This retention pattern may explain the D. acuminata maxima found in the study area
332	during the survey and its later dispersion and disappearance (Fig 3, Fig 8).
333	In contrast, although the minimum-retention periods in the whole BV region (Area 2; Fig
334	11B) coincided with those found in Area 1, the proportion of retained particles in Area 2
335	(35%) was much higher than in the estimations for Area 1 (5%). Therefore, although
336	dispersed in a larger area, the virtual population of Dinophysis still remained in the BV
337	after each simulation, even during high-dispersion periods.
338	The Ichthyop model was also used to simulate the retention-dispersion of the D .
339	acuminata maximum found at station 31 (2-5 m depth; Fig 12). Retention rates decreased
340	after 11 July and the particle population was advected southwardly immediately after
341	their release (Fig 12). These results agree with the distribution of D. acuminata found
342	during the second leg of the cruise.

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4. Discussion

The growing season of *D. acuminata*—defined to occur from April to September in the BV (Xie et al., 2007)—started by the end of April in 2006, but finished unusually early, by the end of July, i.e. one month after reaching its seasonal cell maximum (19 June). Although high concentrations ($>10^3$ cell L⁻¹) of *Dinophysis* spp. were not found during the cruise, substantial information on the conditions associated with the bloom decline and the physiological status of *D. acuminata* was gathered. Division rates of D. acuminata estimated from short-term incubations in culture-plate wells with addition of DOM (> 1 kD) to the local seawater were significantly higher than those from the control-wells with only local seawater. These results suggest a beneficial effect of DOM on *Dinophysis* growth that would explain the frequent co-occurrence of D. acuminata cell maxima with those of organic aggregates (Lunven et al., 2005). Maximum values of μ (0.23 d⁻¹) were 2-4 times lower than those observed under optimum culture conditions of D. acuminata fed on the ciliate Myrionecta rubra (= Mesodinium rubrum) by Park et al., 2006; Riisgaard and Hansen, 2009. These authors concluded that Dinophysis is an obligate mixotroph that requires both light and live prey for long term survival. Nevertheless, the enhanced growth observed in the DOM-enriched incubations suggests that DOM may act as a nutritional supplement that helps to maintain and even to increase population numbers between successive feedings on *Myrionecta*. The high frequency of viable cells, the moderate division rates observed on the on board incubations and the healthy appearance of the cells observed immediately after

365 samples recollection indicates that the D. acuminata population under study was in good 366 physiological conditions. 367 The decline of dinoflagellate blooms has been thought to result from a combination of 368 physical and biological factors (Steidinger, 1973; Garcés et al., 1999). Amongst the 369 biological processes, sexual reproduction and encystment, aging of the population and 370 grazing have been identified as potential key factors to understand the bloom decline 371 (Calbet et al., 2003). Amongst the potential physical factors, changes in wind velocity 372 and direction, turbulence and mixing may also account for the decline of a dinoflagellate 373 population. 374 The importance of water transport in the population dynamics of *Dinophysis* spp. and 375 its possible advection to the coast has been described in several studies (i.e. Belgrano et 376 al., 1999; Delmas et al., 1992). Physical driving forces, such as wind and/or currents that 377 provoke accumulation-dispersion of *Dinophysis* cells have been already reported by some 378 authors (Koike et al., 2001; Koukaras and Nikolaidis, 2004; Soudant et al., 1997). More 379 recently, *Dinophysis* events have been related to the onshore transport of an eddy, located 380 offshore in the BV (Xie et al., 2007), that may constitute a retention area or "incubator" 381 for D. acuminata. Xie et al. (2007) related increased densities of Dinophysis spp. to 382 retentive zones where horizontal dispersion of the growing population was limited. Soudant et al. (1997) explained *D. acuminata* abundance in Antifer (Normandy, France) 383 384 using a dynamic linear regression. They concluded that the disappearance of D. 385 acuminata cells in the area was mainly driven by northeasterly winds that provoked cells 386 dispersion and by increased tidal coefficients that induced D. acuminata dispersal by 387 dilution and water- masses movement. Our results from Ichthyop simulations favour the

388 consideration of these retention-dispersion patterns, since simulations of minimum 389 particle retention coincided with the disappearance of the D. acuminata peaks on 19 June 390 (REPHY data; Fig 4) and on 11 July (Cruise data; Fig 5). Moreover, maximum 391 percentages of particle dispersion were found at the surface, where the D. acuminata cell 392 maxima were observed (Table 2) during the cruise. 393 High dilution rates promoted by sea currents can be expected to return dinoflagellate 394 populations to earlier developmental phases or even disperse them completely (Zingone 395 and Wyatt, 2004). Results from Ichthyop show that although D. acuminata may have 396 been dispersed from its bloom location (Area 1), a percentage (10 %) of the population 397 remained in the whole BV (Area 2). Those cells still present in the area may have acted 398 as the seed for the next D. acuminata peak detected after the cruise by REPHY (Fig 4) 399 before their definitive disappearance at the beginning of August 2006. 400 Biological processes (i.e. plankton behaviour, grazing, growth, and mortality) can be 401 important factors that may influence the outcomes of 3D LPTM. The grazing impact of 402 zooplankton has been identified as a significant loss factor that could affect HABs 403 dynamics (Turner and Anderson, 1983; Watras et al., 1985). Grazing from protist such as 404 heterotrophic dinoflagellates and ciliates has been shown to cause considerably mortality 405 (< 50%) of even collapse of dinoflagellate blooms (Kamiyama et al., 2001; Kamiyama 406 and Matsuyama, 2005; Matsuyama et al., 1999). Following dissection of specimen from 407 samples collected during our cruise, Nézan and Chomérat (2009) observed that cells of 408 the heterotrophic dinoflagellate, Fragilidium duplocampanaeforme, often contained theca 409 of D. acuminata in their interior. The unusual high concentrations of ecdysal cysts of 410 Fragilidium observed during the first leg of the cruise combined with the moderate

411 concentrations of *Dinophysis* may have caused a high grazing impact of the former in the 412 D. acuminata. 413 There are other interactions between organisms, different from predator-prey 414 relationships that may have a strong impact on the population dynamics of harmful algae. 415 In this context, allelopathy has been considered as a potential agent that could promote 416 bloom decline and as a key factor in the control of algal succession. Among others, the 417 allelopathic effects of *Chrysochromulina polylepis* on several plankton species include an 418 initial decrease in growth rate of the tested algae, followed by a decline in their 419 population numbers (Schmidt and Hansen, 2001). In our study, the decline of D. 420 acuminata was observed coinciding with the onset of a Chrysochromulina sp. bloom that 421 formed a conspicuous thin layer (10 m depth) during the second leg of the cruise. 422 Parasitism is also recognized as an important microbial control of bloom-forming 423 dinoflagellates, with species of Amoebophrya being particularly noteworthy, as they are 424 widely distributed in coastal environments and infect numerous host taxa, including 425 several toxin-producing species (Park et al., 2004). Recently, D. acuminata has been 426 reported to be one of the many hosts for Amoebophrya spp. (Salomon et al., 2009; 427 Gonzalez-Gil et al., accepted) and for that, especial attention was paid to detect infected 428 D. acuminata cells during the survey. Although epidemic outbreaks of Amoebophrya spp. 429 have been reported for several dinoflagellate species, such as Alexandrium catenella 430 (Nishitani et al., 1985; Taylor, 1968), Ceratium falcaltiforme (Salomon et al., 2009) and 431 D. norvegica (Salomon et al., 2003), .D. acuminata has not shown high infection levels in 432 field studies (1-3 %, Gonzalez-Gil et al., Personal communication). During this study,

433 frequencies of infected cells were below detection limits and parasites, at least in this 434 situation, can be discarded as a relevant loss factor 435 The relationship between *Dinophysis* and environmental factors has been extensively 436 considered in many studies (reviewed in Maestrini, 1998). The summer 1983, defined in Lassus et al. (1985) as a very hot summer, Dinophysis populations were found between 437 15-21°C in Southern Brittany coasts. The highest abundances during that year were 438 439 observed at 15°C, and corresponded to the beginning of the population dispersal. Besides, 440 although large surface lenses of water of 24°C were detected in southern Brittany in that 441 summer, *Dinophysis* spp. populations were only found within the BV, characterized at 442 that moment by moderate surface temperatures (below 18°C). During the HABIT cruise 443 on July 2006, D. acuminata was mainly found near the surface, where the temperature 444 was significantly higher (19-20.51°C) than the average typical summer conditions in the 445 area (Fig 7). 446 The effect of daily vertical migration (DVM) on retention and dispersion of D. 447 acuminata populations has not been taken into account in our model simulations, and 448 indeed, no signs of DVM of this species were observed during our cruise. However, D. 449 acuminata cells are not passive particles and some evidence has been presented for its 450 DVM under very calmed conditions in Ría de Vigo (Villarino et al., 1995). Nevertheless, in other situations and in the same region, D. acuminata has not been observed to perform 451 452 any vertical migration M (Velo-Suárez et al., 2008; González-Gil et al., accepted). It 453 seems that DVM is not a constant feature in the behaviour of D. acuminata, and further 454 studies are needed to decide under which circumstances virtual particles should be 455 programmed to have DVM in model simulations.

Biological observations and analysis during the survey have emphasized the importance of biological processes in HABs population dynamics. However, the parameterization and calibration of these processes (i.e. mortalities induced by predation, allelopathy, and temperature effects, parasitism and swimming behaviour and buoyancy) are not a simple task and species-specific vital rates are needed to include them into individual based models and obtain accurate simulations of D. acuminata dynamics in the future. The first application of a 3D LPTM with a HAB species shows how these models are able to reproduce well D. acuminata patches and dispersion patterns obtained in the field. Comparisons between survey data (D. acuminata concentrations) and model output (% of virtual particle retention) were in good agreement (Fig 4, Fig 8 and Fig 12), considering that an exact correspondence between model outputs and observations in a system with strong meso- and sub- mesoscale activity is difficult to achieve. Physical factors (windgenerated turbulence, advection, etc) can explain the rapid collapse of a HAB population but biological interactions cannot be dismissed during the decline processes. In order to understand the interactions in the pelagic environment, we need to integrate biology and physics on the temporal and spatial scales of the population dynamics of the individual species. Results from this work give confidence to set up, in a near future, a 3D LPTM of the dynamics of *D. acuminata* for the Northern Bay of Biscay that should include vertical migration patterns and other important biological aspects.

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484	7. References
485	Aminot, A., Kerouel, R., 2004. Hydrologie des écosystèmes marins. Paramètres et
486	analyses. Collection Méthodes d'analyse en milieu marin. IFREMER, France.
487	Belgrano, A., Lindahl, O., Hernroth, B., 1999. North Atlantic Oscillation, primary
488	productivity and toxic phytoplankton in the Gullmar Fjord, Sweden (1985-1996).
489	Proc. Roy. Soc. Lond. B. Bio. 266, 425–430.
490	Benner, R., Biddanda, B., Black, B., McCarthy, M., 1997. Abundance, size distribution,
491	and stable carbon and nitrogen isotopic compositions of marine organic matter
492	isolated by tangential-flow ultrafiltration. Mar. Chem. 57, 243–263.
493	Calbet, A., Vaqué, D., Felipe, J., Vila, M., Sala, M. M., Alcaraz, M., Estrada, M., 2003.
494	Relative grazing impact of microzooplankton and mesozooplankton on a bloom of
495	the toxic dinoflagellate Alexandrium minutum. Mar. Ecol. Prog. Ser., 259, 303–
496	309.
497	Coleman, N. K., Vestal, J. R., 1987. An epifluorescent microscopy study of enzymatic
498	hydrolysis of fluorescein diacetate associated with the ectoplasmic net elements of
499	the protist <i>Thraustochytrium striatum</i> . Can. J. Microbiol. 33, 841–843.
500	Delmas, D., Herbland, A., Maestrini, S. Y., 1992. Environmental conditions which lead

to increase in cell density of the toxic dinoflagellates *Dinophysis* spp. in nutrient-rich and nutrient-poor waters of French Atlantic coast. Mar. Ecol. Prog. Ser. 89, 53–61.

504	Delmas, D., Herbland, A., Maestrini, S. Y., 1993. Do <i>Dinophysis</i> spp. come from the
505	"open sea" along the French Atlantic coast? in: Smayda, T., Shimizu, Y. (Eds.),
506	Toxic phytoplankton blooms in the sea. Elsevier, Amsterdam, pp. 489-494.
507	Dodson, A. N., Thomas, W. H., 1978. Reverse filtration, in: Sournia, A. (Ed.),
508	Phytoplankton Manual. UNESCO, Paris, pp. 104-106.
509	Gallego, A., North, E. W., Petitgas, P., 2007. Introduction: status and future of modelling
510	physical-biological interactions during the early life of fishes. Mar. Ecol. Prog.
511	Ser. 347, 122–126.
512	Garcés, E., Masó, M., Camp, J., 1999. A recurrent and localized dinoflagellate bloom in a
513	Mediterranean beach. J. Plankton Res. 21, 2373–2391.
514	Gentien, P., 1986. A method for evaluating phytoplankton viability by induced
515	fluorochromasia, in: Manzoli, F. A. (Ed.), Progress in flow cytometry. ISBN 90-
516	9001361-X, pp. 151–164.
517	Gentien, P., Lunven, M., Lehatre, M., Dunvent, J. L., 1995. <i>In situ</i> depth profiling of
518	particles sizes. Deep Sea Res. I 42, 1297–1312.
519	González-Gil, S., Velo-Suárez, L., Gentien, P. Ramilo, I., Reguera, R., accepted.
520	Phytoplankton assemblages and characterization of Dinophysis acuminata
521	population during an upwelling-donwelling cycle. Aquat. Microb. Ecol.
522	Guo, L., Coleman, C. H. J., Santschi, P. H., 1994. The distribution of colloidal and
523	dissolved organic carbon in the Gulf of Mexico. Mar. Chem. 45, 105–119.

) 24	Kamiyama, 1., Matsuyama, 1., 2005. Temporal changes in the chiate assemblage and
525	consecutive estimates of their grazing effect during the course of a Heterocapsa
526	circularisquama. J. Plankton Res. 27, 303–311.
527	Kamiyama, T., Takayama, H., Nishii, Y., Uchida, Y., 2001. Grazing impact of the field
528	ciliate assemblage on the bloom of the toxic dinoflagellate Heterocapsa
529	circularisquama. Plankton Biol. Ecol. 48, 10–18.
530	Kim, S., Kang, Y. G., Kim, H. S., Yih, W., Coats, D. W., Park, M. G., 2008. Growth and
531	grazing responses of the mixotrophic dinoflagellate Dinophysis acuminata as
532	functions of light intensity and prey concentration. Aquat. Microb. Ecol. 51, 301-
533	310.
534	Koike, K., Otobe, H., Takagi, M., Yoshida, T., Ogata, T., Ishimaru, T., 2001. Recent
535	occurrences of Dinophysis fortii (dinophyceae) in the Okkirai Bay, Sanriku,
536	Northern Japan, and related environmental factors. J. Oceanogr. 57, 165–175.
537	Koukaras, K., Nikolaidis, G., 2004. Dinophysis blooms in the Greek coastal waters
538	(Thermaikos Gulf, NW Aegean Sea). J. Plankton Res. 26, 445–457.
539	Lassus, P., Barduouil, M., Trunquet, I., Trunquet, P., Le Baut, C., Pierre, M. J., 1985.
540	Dinophysis acuminata distribution and toxicity along the Southern Brittany Coast
541	(France): correlation with hydrological parameters, in: Anderson, D. M., White,
542	A. W., Baden, D. G. (Eds.), Toxic Dinoflagellates. Elsevier Science Publishing,
543	Inc, pp. 159–164.

)44	Lazure, P., Dumas, F., 2008. An external-internal mode coupling for a 3D
545	hydrodynamical model at regional scale (MARS). Adv. Wat. Res. 31, 233–250.
546	Lazure, P., Dumas, F., Vrignaud, C., 2008. Circulation on the armorican shelf (Bay of
547	Biscay) in autumn. J. Marine Syst. 72, 218–237.
548	Lazure, P., Jegou, A. M., 1998. 3D modelling of seasonal evolution of Loire and Gironde
549	plumes on Biscay Bay continental shelf. Oceanol. Acta 21, 165–177.
550	Lett, C., Verley, P., Mullon, C., Parada, C., Brochier, T., Penven, P., Blanke, B., 2008. A
551	Lagrangian tool for modelling ichthyoplankton dynamics. Environ. Modell.
552	Softw. 23, 1210–1214.
553	Lunven, M., Gentien, P., Kononen, K., Le Gall, E., Danielou, M. M., 2003. <i>In situ</i> video
554	and diffraction analysis of marine particles. Estuar. Coast. Shelf Sci. 57, 1127-
555	1137.
556	Lunven, M., Guillaud, J. F., Youenou, A., Crassous, M. P., Berric, R., Le Gall, E.,
557	Kerouel, R., Labry, C., Aminot, A., 2005. Nutrient and phytoplankton distribution
558	in the Loire River plume (Bay of Biscay, France) resolved by a new Fine Scale
559	Sampler. Estuar. Coast. Shelf Sci. 65, 94–108.
560	Maestrini, S. Y., 1998. Bloom dynamics and ecophysiology of <i>Dinophysis</i> spp, in:
561	Anderson, D. M., Cembella, A. D., Hallegraeff, G. M. (Eds.), Physiological
562	ecology of harmful algae blooms. Vol. NATO ASI, Series G of Ecological
563	Sciences 41. Springer, Berlin Heidelberg, New York, pp. 243–266.

004	Maestrini, S. Y., Berland, B. R., Grzebyk, D., Spano, A. M., 1995. <i>Dinophysis</i> spp. cells
565	concentrated from nature for experimental purposes, using size fractionation and
566	reverse migration. Aquat. Microb. Ecol. 9, 177–182.
667	Marcaillou, C., Gentien, P., Lunven, M., Grand, J. L., Mondeguer, F., Daniélou, M. M.,
568	Crassous, M., Youenou, P., 2001. Dinophysis acuminata distribution and specific
69	toxin content in relation to mussel contamination, in: Hallegraeff, G. M.,
570	Blackburn, S. I., Bolch, C. J., Lewis, R. J. (Eds.), Harmful Algal Blooms. IOC of
571	UNESCO, Paris, pp. 356–359.
572	Matsuyama, Y., Miyamoto, M., Kotani, Y., 1999. Grazing impacts of the heterotrophic
573	dinoflagellate Polykrikos kofoidii on a bloom of Gymnodinium catenatum. Aquat
574	Microb. Ecol. 17, 91–98.
575	Nishitani, L., Erickson, G., Chew, K. K., 1985. Role of the parasitic dinoflagellate
576	Amoebophrya ceratii in control of Gonyaulax catenella populations, in:
577	Anderson, D. M., White, A. W., Baden, D. G. (Eds.), Toxic dinoflagellates.
578	Elsevier, New York, pp. 225–230.
579	Nézan, E., Chomérat, N., 2009. Fragilidium duplocampanaeforme sp. nov.
580	(Dinophyceae): A new phagotrophic dinoflagellate from the French Atlantic
581	coast. Eur. J. Protistol 45, 2–12.
582	Park, M. G., Kim, S., Kim, H. S., Myung, G., Kang, Y. G., Yih, W., 2006. First
883	successful culture of the marine dinoflagellate Dinophysis acuminata in cultures.
584	Aquat. Microb. Ecol. 45, 101–106.

085	Park, M. G., Yin, W., Coats, D. W., 2004. Parasites and phytopiankton, with special
586	emphasis on dinoflagellate infections. J. Eukaryot. Microbiol. 51, 146–155.
587	Purina, I., Balode, M., Béchemin, C., Põder, T., Verité, C., Maestrini, S., 2004. Influence
588	of dissolved organic matter from terrestrial origin on the chances of dinoflagellate
589	species composition in the Gulf of Riga, Baltic Sea. Hydrobiologia 514, 127-137.
590	Riisgaard, K., Hansen, P. J., 2009. Role of food uptake for photosynthesis, growth and
591	survival of the mixotrophic dinoflagellate Dinophysis acuminata. Mar. Ecol. Prog
592	Ser. 381, 51–62.
593	Salomon, P. S., Graneli, E., Neves, M. H. C. B., Rodríguez, E. G., 2009. Infection by
594	Amoebophrya spp. parasitoids of dinoflagellates in a tropical marine coastal area.
595	Aquat. Microb. Ecol. 55 (2), 143–153.
596	Salomon, P. S., Janson, S., Graneli, E., 2003. Parasitism of <i>Dinophysis norvegica</i> by
597	Amoebophrya sp in the Baltic Sea. Aquat. Microb. Ecol. 33 (2), 163–172.
598	Santos, A. M. P., Chicharo, A., Dos Santos, A., Moita, T., Oliveira, P. B., Peliz, A., Re,
599	P., 2007. Physical-biological interactions in the life history of small pelagic fish in
500	the western iberia upwelling ecosystem. Prog. Oceanogr. 74 (2-3), 192–209.
501	Schmidt, L. E., Hansen, P. J., 2001. Allelopathy in the prymnesiophyte
502	Chrysochromulina polylepis: Effect of cell concentration, growth phase and pH.
503	Mar. Ecol. Prog. Ser. 216, 67–81.

604	Soudant, D., Beliaeff, B., Thomas, G., 1997. Explaining <i>Dinophysis</i> cf. acuminata
605	abundance in Antifer (Normandy, France) using dynamic linear regression. Mar.
606	Ecol. Prog. Ser. 156, 67–74.
607	Steidinger, K. A., 1973. Phytoplankton ecology: a conceptual review based on eastern
608	Gulf of Mexico research. CRC Rev. Microb. 3, 49–68.
609	Taylor, F. J. R., 1968. Parasitism of the toxin-producing dinoflagellate <i>Gonyaulax</i>
610	catenella by the endoparasitic dinoflagellate Amoebophrya ceratii. J. Fish. Res.
611	Board. Can. 25, 2241–2245.
612	Turner, J. T., Anderson, D. M., 1983. Zooplankton grazing during dinoflagellate blooms
613	in a Cape Cod embayment, with observations of predation upon tintinnids by
614	copepods. Mar. Ecol. 4, 359–374.
615	Utermöhl, H., 1931. Neue wege in der quantitativen Erfassung des Planktons (mit
616	besonderer Berücksichtigung des Ultraplanktons). Verh. Int. Ver. Theor. Angew.
617	Limnol. 5, 567–596.
618	Velo-Suárez, L., González-Gil, S., Gentien, P., Lunven, M., Bechemin, C., Fernand, L.,
619	Raine, R., Reguera, B., 2008. Thin layers of <i>Pseudo-nitzschia</i> spp. and the fate of
620	Dinophysis acuminata during an upwelling-downwelling cycle in a Galician Ría.
621	Limnol. Oceanogr. 53, 1816–1834.
622	Velo-Suárez, L., Reguera, B., Garces, E., Wyatt, T., 2009. Vertical distribution of
623	division rates in coastal dinoflagellate Dinophysis spp. populations: implications
624	for modelling. Mar. Ecol. Prog. Ser. 385, 87–96.

625	Villarino, M. L., Figueiras, F. G., Jones, K. J., Álvarez-Salgado, X. A., Richard, J.,
626	Edwards, A., 1995. Evidence of in situ diel vertical migration of a red-tide
627	microplankton species in Ría de Vigo (NW Spain). Mar. Biol. 123, 607-617.
628	Watras, C. J., Garcon, V. C., Olson, R. J., Chisholm, S. W., Anderson, D. M., 1985. The
629	effect of zooplankton grazing on estuarine bloom of the toxic dinoflagellate
630	Gonyaulax tamerensis. J. Plankton Res. 7, 891–908.
631	Xie, H., Lazure, P., Gentien, P., 2007. Small scale retentive structures and <i>Dinophysis</i> . J.
632	Marine Syst. 64, 173–188.
633	Yasumoto, T., Murata, M., Oshima, Y., Sano, M., Matsumoto, G. K., Clardy, J., 1985.
634	Diarrhetic shellfish toxins. Tetrahedron 41 (6), 1019–1025.
635	Zingone, A., Wyatt, T., 2004. Harmful Algal Blooms: Keys to the understanding of
636	phytoplankton ecology, in: Robinson, A. R., Brink, K. H. (Eds.), The Global
637	Coastal Ocean: Multi-Scale Interdisciplinary Processes. The Sea, Volume 13.
638	Harvard University Press, pp. 867–926.
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640 Fig 1: (A) Map of Western Europe and location of the Northern Bay of Biscay (B) 641 Location of the Bay of Vilaine (BV) in the Northern Bay of Biscay. 642 643 Fig 2: Spatial distribution of sampling sites in the Northern Bay of Biscay during the first 644 (A) and the second leg (B) of the HABIT 2006 survey. 645 646 Fig 3: Weekly distribution of *Dinophysis* spp. for 2006 in the BV (Data from REPHY). 647 The extension of the *Dinophysis* season is highlighted with an arrow and the period when 648 HABIT2006 cruise was carried out is marked as a shadowed area. 649 650 Fig 4: (A) Horizontal distribution of *D. acuminata* cell maxima in the Northern Bay of Biscay during the first leg of the survey (06-12 July 2006). (B) Detail of the D. 651 652 acuminata horizontal abundance in the BV (06-12 July 2006). 653 654 Fig. 5: Horizontal distribution of (A) salinity, (C) temperature and (E) particle load 655 measured with the IPSAP profiler in the study area from 06-12 July. Triangles in A, C 656 and E indicate sampling points during the first leg. Vertical distribution of (B) salinity, 657 (D) temperature and (F) particle load along a transect off the Loire estuary (shown as a line in panel A). D. acuminata maxima (cell L⁻¹) and their location in the water column 658 659 are also marked in panels B, D, and F. 660 Fig. 6: Distribution of sea surface temperature (SST) anomalies at the Northern Bay of 661 Biscay on 12 July 2006 (EUMETSAT). Average SST was obtained from a 22-yr time 662 series (1986-2008).

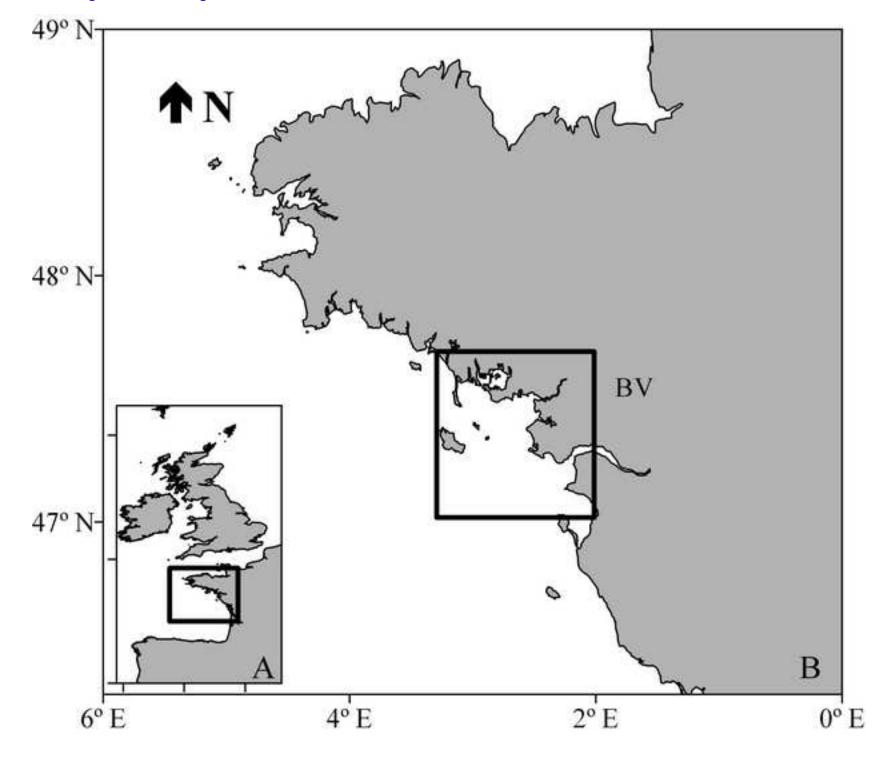
663 Fig. 7: Vertical distribution of *D. acuminata* at station 31 (BV) on 11 July 2006. 664 Fig. 8: (A) Horizontal distribution of D. acuminata cell maxima in the Northern Bay of 665 666 Biscay during the second leg of the survey (13-22 July 2006). (B) Detail of the horizontal 667 distribution of *D. acuminata* in the BV (13-22 July 2006). 668 669 Fig. 9: An FDA-treated field sample. (A) The black arrow indicates metabolically-active 670 cell of D. acuminata in DIC micrograph, and (B); yellow arrow indicate non 671 metabolically-active cells with weak fluorescence or with no fluorescence at all. 672 Fig. 10: Incubations (5 cell well⁻¹) of *D. acuminata* with and without enrichment with 673 DOM>1 kD from BV. Moderate growth: μ_{max} = 0.23 d⁻¹ (0.33 div d⁻¹) was observed in 674 675 DOM-enriched wells. 676 677 Fig. 11 Percentages of retained particles obtained from Ichthyop runs in the BV from 15 678 June to 30 July 2006 in (A) Area 1 (sta 31) and (B) the whole BV. Plot shaded areas mark 679 when D. acuminata reached its maximum concentration (data on 19 July from REPHY 680 and on 11 July from cruise) in the BV. 681 Fig. 12 Results from Ichtyop transport scenarios from 09 July to 19 July at 2-5 m depth. 682 683 (A) Map of the Northern Bay of Biscay showing Area 2. Position of particles after (B) 2day, (C) 4-day, (D) 6-day, (E) 8-day, and (F) 10-day simulation are shown as black dots. 684 685 Red box indicates the location of Area 1.

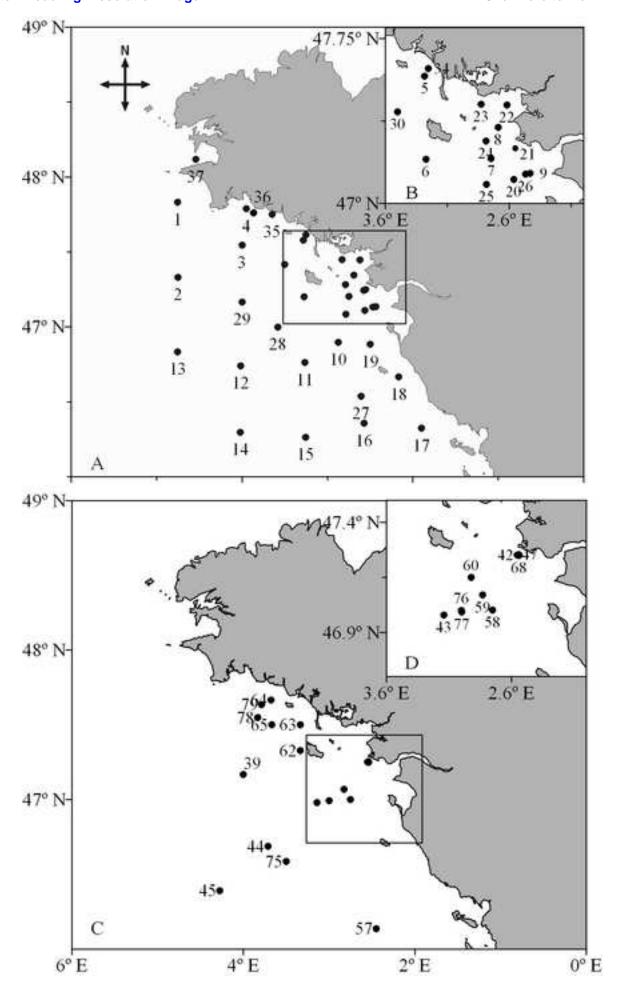
Table 1: Temporal distribution of sampling sites in the Northern Bay of Biscay during the first and the second leg of the HABIT 2006 survey.

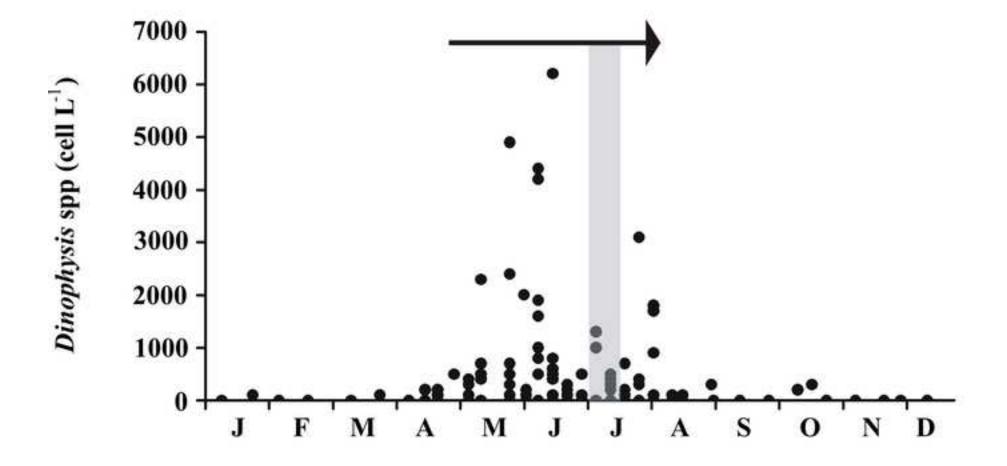
	Leg 1	Leg 2		
Date (July)	Sampling stations	Date (July)	Sampling stations	
06	1-4	13	38	
07	5-10	14	39-41	
08	11-15	15	42-45	
09	16-21	16	47	
10	22-28	18	58-60,62-63	
11	29-33	19	64-68	
12	34-37	20	71	
		21	75,77	
		22	78-79	

Table 2: Depth of *D. acuminata* cell maxima at each station where its concentration in the whole water column reached 100 cells L^{-1} during the first leg of the cruise (06-12 July).

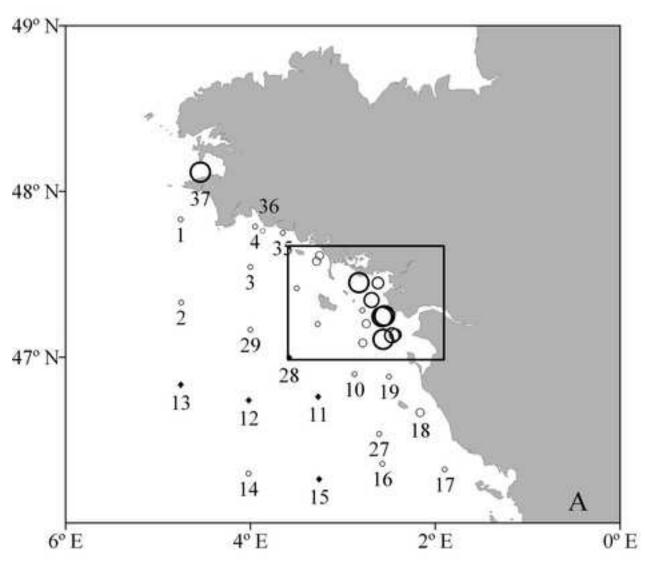
Station	Day	Sampling hour	D. acuminata	Depth	
		(GMT)	cells L ⁻¹	(m)	
8	07 July	12:30	304	7	_
9	07 July	15:30	148	1.7	
20	09 July	21:00	933	4	
21	09 July	23:30	707	4	
23	10 July	04:00	603	16	
26	10 July	01:00	487	1	
31	11 July	12:30	1002	2.25	
37	12 July	12:00	632	2	

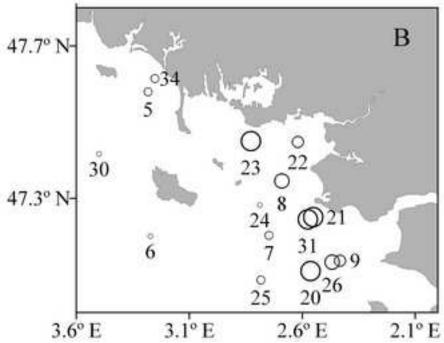


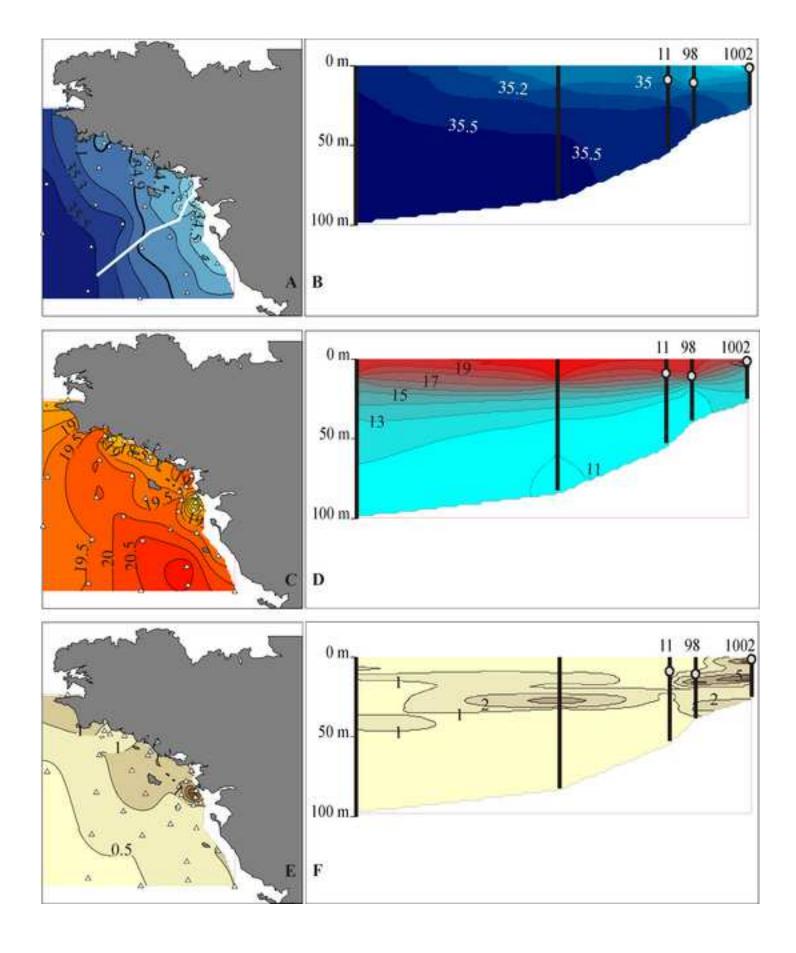


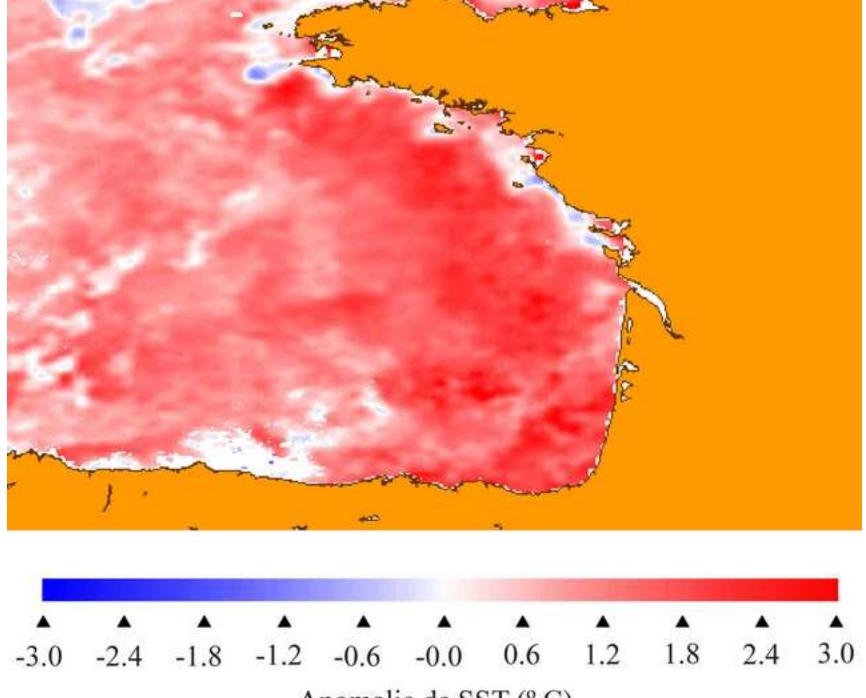


* not detected ○ <50 ○ 50 - 100 ○ 100-250 ○ 250-500 ○ 500-1000 cell L-1









Anomalie de SST (° C)

