

Approaches to Model the Life Cycle of Harmful Algae

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Abstract

Models of harmful algal blooms (HABs) need to include autecological characteristics of the HAB species because the bloom dynamics can only be successfully described if certain life cycle aspects (in particular en- and excystment) are included in some way. In this study, an overview is presented on how the life cycle is considered in current Lagrangian and Eulerian models. Examples of the latter are given, which range from crude parameterizations in one- compartment models, to stage-resolving twelve-compartment models. Advantages and disadvantages of the different approaches are highlighted. A generalized model classification is presented which may be used as a framework for further phytoplankton life cycle modeling studies.

Key words: life cycle, model, phytoplankton, harmful algal blooms (HABs), seed population, concept

1. Introduction

Harmful algal blooms (HABs; here defined as "high" biomass but not necessarily toxic blooms), are frequently observed in coastal areas but their causes are still somewhat obscure. Many different species of microalgae with their different requirements for optimal growth can form HABs. Some HABs seem to occur entirely naturally, as part of the seasonal succession of marine organisms, others

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7 seem to be triggered by environmental changes introduced by human activities
8 (see, e.g., Anderson et al., 2002).

9 HABs typically appear – and disappear – quite abruptly. This is surpris-
10 ing because the growth rates of many HAB forming species are comparatively
11 low (see, e.g. Stolte and Garcés, 2006). Similarly, blooms are observed to end
12 even though the environmental conditions that are considered favorable persist
13 (Anderson et al., 1983; Kremp and Heiskanen, 1999). This apparent paradox
14 can be explained as the result of the species life cycle. Transitions between veg-
15 etative and resting phases – the formation of resting stages (encystment) and
16 the reverse process excystment (germination) – are responsible for terminating
17 or initiating a bloom (e.g., Anderson, 1998; Garcés et al., 2002; Anderson and
18 Rengefors, 2006). Resting cells from a previous bloom may have settled on
19 the bottom where they may become an additional benthic source of biomass,
20 when germinating simultaneously and rising in synchrony with the onset of the
21 bloom of the pelagic population. The number of excysting resting cells recruit-
22 ing the water column may actually be the most important factor to determine
23 the magnitude of the bloom.

24 The various life history strategies of different HAB species do not only in-
25 fluence timing, magnitude and duration of blooms but also the dominance and
26 seasonal succession of species (Anderson and Rengefors, 2006; Kremp et al.,
27 2008). For example, different cyst-forming dinoflagellate species have different
28 temperature windows for germination explaining the seasonal succession of the
29 respective populations (Anderson and Rengefors, 2006).

30 Large year-to-year fluctuations in the abundance of harmful algae are ob-
31 served, even though the environmental conditions (light, temperature, nutrients)
32 vary much less. This can be explained either by a very high sensitivity to minute
33 details in these parameters, by the influence of an unknown external quantity,
34 or by variations in the inoculum (Kremp et al., 2008) as a result of the life cycle
35 aspects (e.g., maturation times).

36 Although the life cycle appears to be essential, it has long been ignored in
37 ecosystem models with HAB species. The reason may be related to the com-

plexity of the life cycle (Fig. 1) of which we often possess only rudimentary knowledge. First of all, it is unclear how many distinct stages need to be considered. The triggering factors leading to life cycle stage transition are not fully understood. Finally, for each life cycle stage we need to specify parameters with respect to nutrient demands (e.g., uptake of dissolved inorganic nitrogen versus nitrogen fixation), motility (e.g., sinking versus rising), tolerance for physical factors (e.g., salinity, temperature, light, turbulence), mortality, respiration and growth.

This paper presents an overview of existing strategies to include life cycle aspects in numerical ecosystem models (in the following LCM: Life Cycle Models), in particular for species with growing and resting stages. The advantages and disadvantages are discussed and future directions are outlined.

2. The Lagrangian Approach

The most natural way to consider the life cycle of an organism is certainly to use an individual-based (or agent-based) approach, following the organism in time through the fluid (advection) and through the different stages of its development (life cycle succession). In other words, organisms are treated as discrete individuals with certain properties.

2.1. Governing Equations

A Lagrangian model for a population of identical organisms involves a set of four equations:

$$\frac{\partial P}{\partial t} = \mu P - l_P P \quad (1)$$

$$\frac{\partial x^P}{\partial t} = u \quad (2)$$

$$\frac{\partial y^P}{\partial t} = v \quad (3)$$

$$\frac{\partial z^P}{\partial t} = w + w_P \quad (4)$$

where P is the phytoplankton biomass, t the time, μ the actual growth rate, l the loss rate including natural mortality, lysis due to viral infection and grazing

by zooplankton and higher trophic levels. The position of the population in three-dimensional space is given by $\vec{x}^P = (x^P, y^P, z^P)$ and changes according to local fluid velocities $\vec{v} = (u(x, y, z, t), v(x, y, z, t), w(x, y, z, t))$. Vertical motility of the population can be included through w_P .

In a simple model, the actual growth rate will depend on external factors like temperature, light and nutrient availability $\mu = \mu(I, N, T)$, while the loss rate is a function of grazer and virus concentration. Additional dependencies can be included with relative ease.

2.2. Life Cycle Processes

Life cycle related changes and transformations of phytoplankton take place during the development of the organism. In the Lagrangian approach both endogenous and exogenous triggering factors can be easily included. An endogenous clock can be considered by integrating an equation for the "age" (A^P) of the population relative to some reference date:

$$\frac{\partial A^P}{\partial t} = 1 \quad (5)$$

Growth, mortality and migration can then be made a function of age, varying either continuously or abruptly. A maturation time or a mandatory dormancy period of a resting stage may also be considered by prescribing specific times or time periods. Hence, a characteristic c of the population is a function of external factors like irradiance (I), temperature (T), salinity (S), nutrient concentrations (N) and age (A^P):

$$c = c(I, T, S, N, A^P). \quad (6)$$

Mechanistic approaches to phytoplankton dynamics emphasize the role of internal factors for the development of organisms. Cell size, as well as the intracellular availability of energy, nutrients, toxins are probably better descriptors of phytoplankton properties than external concentrations. This means that for N internal properties $\vec{Q}^P = (Q_1^P, Q_2^P, \dots, Q_N^P)$ the governing equations are augmented with another set of equations:

$$\frac{\partial \vec{Q}^P}{\partial t} = \vec{\nu}^P - \vec{\varphi}^P - l\vec{Q}^P \quad (7)$$

where $\bar{\nu}P$ and $\bar{\varphi}P$ are the source and sink terms of the internal pools (which in turn depend on external factors and/or the ratios of internal concentration to total biomass (the quotas $\vec{\theta} = \vec{Q}^P/P$). Instead of irradiance and external nutrient concentration in equation 6, the internal energy and nutrient quotas ($\vec{\theta} = (\theta_1, \theta_2, \dots, \theta_N)$) may be used to determine specific characteristics:

$$c = c(T, S, \vec{\theta}, A^P). \quad (8)$$

2.3. Advantages and Disadvantages

The advantage of the Lagrangian approach is that populations retain a fixed identity as they move with the fluid. Each population can be unique in its properties; the natural diversity can therefore be easily included and life cycle transitions (e.g., a change in maximum specific growth rate with size) can be represented by linking this property to the age or internal state of the organism.

The obvious disadvantages of Lagrangian modeling are that (i) a very large number of particles needs to be considered for an adequate coverage of any three-dimensional model domain, (ii) the representation of diffusion and vertical convection is not straightforward but requires additional assumptions, (iii) the technical overhead for treating the splitting of populations and re-initialization of extinct populations is nontrivial.

2.4. Examples of Lagrangian LCM

Recently, Lagrangian LCMs of harmful algae have been applied to local sites to study the dispersal of a dinoflagellate (Villanoy et al., 2006) or the effect of nutrient reduction on bloom formation of a cyanobacterium (Hellweger et al., 2008).

Villanoy et al. (2006) prescribe the observed cyst distribution on the sea floor as initial values. If a threshold value of bottom velocity is exceeded, the resuspended cysts from the sediment are transformed into the vegetative stage (excystment). The vegetative cells can grow while they are transported with the current field. Encystment is assumed to occur after a specified time period. The model focuses on one bloom period, hence the fate of encysted cells is not

considered. The model was successful in representing the spatial distribution patterns of the bloom in Manila Bay.

In a more sophisticated approach, Hellweger et al. (2008) distinguish five life cycle stages comprised of three vegetative stages and two resting stages. Different characteristics (photosynthesis, respiration, division, phosphorus uptake and vertical velocities) are specified to each stage. In addition, the transition between the individual stages is assumed to be a function of internal factors (e.g., maturation time, cell size). In a one-dimensional model application this LCM could very well represent the seasonal cycle of growing and resting stages.

In these examples the transition between growing and resting stages is a function of either internal or external factors. In a somewhat simpler approach, Woods (2005) sets the timing of en- and excystment at specific days of the year. Overall, these model studies show that Lagrangian methods are well suited to study aspects of bloom formation of cyst forming species in real world applications.

3. The Eulerian Approach

The alternative to the Lagrangian Model approach is to treat a collection of organisms or populations as a continuum and to assign a biomass concentration value at each grid point of the model. This is the traditional way to design for example one-, two-, or three-dimensional NPZD models.

3.1. Governing Equations

In this case, the evolution equation for phytoplankton reads

$$\frac{\partial P}{\partial t} = \underbrace{-\vec{v} \cdot \nabla P + \nabla(\vec{\kappa} \nabla P)}_{\left. \frac{\partial P}{\partial t} \right|_{\text{PHYS}}} + \mu P - lP - w_P \frac{\partial P}{\partial z} \quad (9)$$

where $\vec{v} \cdot \nabla P$ is the advection term and $\nabla(\vec{\kappa} \nabla P)$ represents the turbulent diffusion (with the turbulent diffusivity coefficient $\vec{\kappa}(x, y, z, t)$). Advection and turbulent diffusion are combined into the physical tendency term $\left. \frac{\partial P}{\partial t} \right|_{\text{PHYS}}$. The notation of the biological variables are the same as above.

141 3.2. Life Cycle Processes

142 In an Eulerian model the age information, the average size of the organisms,
 143 or any other internal property can only be included through a set of additional
 144 equations. Again, let N represent the number of internal pools of phytoplankton
 145 then the evolution equation for the state vector $\vec{Q}^P = (Q_1^P, Q_2^P, \dots, Q_N^P)$ reads

$$\frac{\partial \vec{Q}^P}{\partial t} = \frac{\partial \vec{Q}^P}{\partial t} \Big|_{\text{PHYS}} + \vec{\nu}P - \vec{\varphi}P - l\vec{Q}^P - w_P \frac{\partial \vec{Q}^P}{\partial z}, \quad (10)$$

146 where $\vec{\nu}P$ and $\vec{\varphi}P$ are again the source and sink terms of the internal pools or
 147 properties.

148 A particular problem is the diffusion term because mixing assumes that all
 149 elements of a compartment are identical. Hence, mixing of populations with dif-
 150 ferent internal quotas will lead to averaged (i.e., erroneous) internal properties.
 151 A convenient solution is the introduction of so called "subcompartments" (e.g.,
 152 Janowitz and Kamykowski, 1999; Beckmann and Hense, 2004). They represent
 153 distinct parameter ranges for internal characteristics which can be identified as
 154 individual life cycle stages (Hense and Beckmann, 2006). The remaining task is
 155 then to define proper transfer conditions and rates between these subcompart-
 156 ments.

157 For considering the different life cycle stages, the phytoplankton compart-
 158 ment needs to be divided into M subcompartments. The evolution equation for
 159 the state vector $\vec{P} = (P_1, P_2, \dots, P_M)$ then reads

$$\frac{\partial \vec{P}}{\partial t} = \frac{\partial \vec{P}}{\partial t} \Big|_{\text{PHYS}} + \vec{\mu}\vec{P} - \vec{l}\vec{P} - \vec{w}_{\vec{P}} \frac{\partial \vec{P}}{\partial z} + \mathbb{T}\vec{P} \quad (11)$$

160 where $\vec{\mu}$, \vec{l} and $\vec{w}_{\vec{P}}$ are vectors of actual growth rate, loss rate and buoyancy
 161 velocities for each stage, and $\mathbb{T} = \tau_{i,j}$ is the transfer rate matrix between the
 162 individual stages. In principle, the transfer between any two life cycle stages can
 163 take place, so the matrix may be dense. A closed single loop life cycle, however,
 164 in a general multi-compartment Eulerian model is represented by a sparse life

165 cycle succession matrix

$$\mathbb{T} = \begin{pmatrix} -\tau_{1,2} & 0 & \dots & 0 & \tau_{N,1} \\ \tau_{1,2} & -\tau_{2,3} & 0 & \dots & 0 \\ 0 & \tau_{2,3} & \dots & 0 & \dots \\ \dots & 0 & \dots & -\tau_{N-1,N} & 0 \\ 0 & \dots & 0 & \tau_{N-1,N} & -\tau_{N,1} \end{pmatrix} \quad (12)$$

166 where the elements $\tau_{i,j}$ denote the transfer from stage i to stage j .

167 The transfer rates may be specified as a function of external (environmental)
168 factors (irradiance, temperature, salinity, nutrients) only

$$\tau_{i,j} = \tau_{i,j}(I, T, S, N). \quad (13)$$

169 Like for the Lagrangian approach, the functional dependence may be modified
170 to rely on internal factors as well

$$\tau_{i,j} = \tau_{i,j}(T, S, \vec{\theta}). \quad (14)$$

171 3.3. Advantages and Disadvantages

172 Eulerian models share the advantages of all grid point models: a regular
173 resolution of the domain under consideration, and the possibility to compute
174 integral quantities and gradients in a straightforward way. In addition, the
175 effects of subgridscale processes (like turbulence) on the biological variables are
176 treated as for the physical variables.

177 A significant disadvantage is that explicit time information (e.g., of a manda-
178 tory dormancy period) cannot be included in the Eulerian approach. Time scales
179 specified for a transfer between compartments merely represent the time after
180 which the source concentration is reduced by a factor of e . As a result, the
181 specification of transfer rates has to rely on ad hoc choices (see, e.g., Beckmann
182 and Hense, 2004).

183 It should also be noted that the introduction of subcompartments and inter-
184 nal quotas will lead to a significant increase in the number of Eulerian tracers
185 and hence computer resources (memory and computing time).

186 3.4. Examples of Eulerian LCM

187 Recent Eulerian models that include phytoplanktonic life cycle dynamics
 188 focus on cyanobacteria and dinoflagellates. They range from very simple ap-
 189 proaches based on just one compartment to comparatively complex representa-
 190 tions of four different stages and two internal quotas.

191 3.4.1. One-compartment LCM

192 Models that attempt to represent a species with a pronounced life cycle with
 193 only one compartment have to rely heavily on parameterization. As pointed out
 194 in the Introduction, the most important life cycle aspect is the germination of
 195 resting cells in spring. This "seed population" can be represented, very crudely,
 196 by a minimum concentration or a minimum production throughout the year
 197 (Fig. 2A). Technically, this is realized by prescription of a "minimum value"¹.

198 The minimum concentration approach (Kiirikki et al., 2001) reads

$$\frac{\partial P}{\partial t} = \frac{\partial P}{\partial t}\Big|_{\text{PHYS}} + \mu P - l(P - P_0) - w_P \frac{\partial P}{\partial z}, \quad (15)$$

199 with similar notation as above; P_0 is the constant seed population.

200 The minimum production approach (Burchard et al., 2006, applying the
 201 model of Neumann et al. (2002)) reads

$$\frac{\partial P}{\partial t} = \frac{\partial P}{\partial t}\Big|_{\text{PHYS}} + \mu(P + P_0) - lP - w_P \frac{\partial P}{\partial z}. \quad (16)$$

202 The minimum value in both these cases ensures a minimum phytoplankton
 203 concentration in the pre-bloom phase. Thus, as desired, the bloom formation
 204 can take place relatively rapidly, despite low species specific growth rates (see
 205 discussion in Hense and Burchard, 2009). Both approaches yield very similar
 206 results, depending of course on the specific choice of P_0 ; if the same minimum
 207 value is used, the former approach leads to higher primary production, i.e.
 208 nitrogen fixation in case of cyanobacteria (Hense and Burchard, 2009).

¹It should be noted that the use of such a minimum value has often been regarded as a mere numerical necessity rather than a crude parameterization of the life cycle. Therefore, there is a tendency to "forget" this measure in the model description.

209 The main advantage of the use of a minimum value is that it is easy to im-
 210 plement and relatively inexpensive to compute. The most obvious disadvantage
 211 is, however, that part of the interannual variability is artificially suppressed,
 212 because the starting basis for growth is the same each year. This can in prin-
 213 ciple be remedied by varying the minimum value P_0 with time to account for
 214 year-to-year fluctuations (see also next section).

215 3.4.2. 1.5-compartment LCM

216 One way to more realistically represent the spatial and temporal distribution
 217 of the seed population is to add a separate compartment P_2 , which is filled with
 218 an (observed) concentration of the seed population (Fig. 2B). A one-way transfer
 219 from this fixed pool of biomass to the vegetative stage P_1 will then lead to the
 220 desired increase in biomass at rates larger than the maximum specific growth
 221 rate.

222 In our classification of life cycle models, such an approach is called a 1.5-
 223 compartment LCM, because there is no two-way exchange between the com-
 224 partments. The corresponding equations read

$$\frac{\partial P_1}{\partial t} = \frac{\partial P_1}{\partial t} \Big|_{\text{PHYS}} + \mu P_1 - lP + \tau_{2,1}P_2 - w_{P_1} \frac{\partial P_1}{\partial z} \quad (17)$$

$$\frac{\partial P_2}{\partial t} = -\tau_{2,1}P_2, \quad (18)$$

225 where $\tau_{2,1}$ is again the actual transfer rate of biomass (see above). The trans-
 226 fer is occasionally treated as a prescribed (but time-dependent) flux of biomass
 227 through the lower boundary of the model (e.g., Eilertsen and Wyatt, 2000), or
 228 as piecewise constant with varying values for day and night (Yamamoto et al.,
 229 2002). McGillicuddy et al. (2005) consider external factors (temperature, irra-
 230 diance) as well as a prescribed time dependent "germination potential" (which
 231 can be seen as a measure of internal maturation of the cells).

232 Several models incorporate encystment ϵ indirectly as a loss term of phy-
 233 toplankton l (see above): $l = \epsilon + m$, with m being the mortality. Again, the
 234 encystment rate has been determined to depend on internal factors, e.g., phos-
 235 phorus quota (e.g., Yamamoto et al., 2002) or as a function of external factors

236 using a measure of nutrient limitation (McGillicuddy et al., 2005).

237 The advantage of this class of models is that a realistic spatially and tempo-
 238 rally variable seed population can be taken into account. Even if the knowledge
 239 about the actual transfer rate are sparse, the model will cover the spatial and
 240 temporal variability of the vegetative stage much better than assuming no or a
 241 constant seed population (see, e.g., McGillicuddy et al., 2005). However, such
 242 an approach is restricted to locations and time where information about these
 243 resting cysts is available and can be used as a "boundary" condition for the
 244 model.

245 3.4.3. A two-compartment LCM

246 A two-compartment LCM is the most simple version of a model that explic-
 247 itly resolves a fully closed life cycle with a two-way transfer (Fig. 2C). The cycle
 248 succession matrix then becomes

$$\tau_{2 \times 2} = \begin{pmatrix} -\tau_{1,2} & \tau_{2,1} \\ \tau_{1,2} & -\tau_{2,1} \end{pmatrix} \quad (19)$$

249 Assuming that we distinguish between the growing stage P_1 and the resting
 250 stage P_2 the corresponding equations read:

$$\frac{\partial P_1}{\partial t} = \frac{\partial P_1}{\partial t} \Big|_{\text{PHYS}} + \mu_1 P_1 - l_1 P_1 + \tau_{2,1} P_2 - \tau_{1,2} P_1 - w_{P_1} \frac{\partial P_1}{\partial z} \quad (20)$$

$$\frac{\partial P_2}{\partial t} = \frac{\partial P_2}{\partial t} \Big|_{\text{PHYS}} + \mu_2 P_2 - l_2 P_2 - \tau_{2,1} P_2 + \tau_{1,2} P_1 - w_{P_2} \frac{\partial P_2}{\partial z} \quad (21)$$

251 where the growth rate μ_2 is much smaller (or even zero) than μ_1 , and the
 252 mortality rate l_2 is much smaller than l_1 . Without additional equations for
 253 internal quotas, the transfer between these two stages has to be specified as a
 254 function of external factors (see above).

255 The corresponding vertical velocities can be chosen in various ways. An ob-
 256 vious choice is to assign a small positive or neutral buoyancy to the vegetative
 257 stage w_{P_1} while w_{P_2} represents sinking (after encystment) and rising (after ex-
 258 cystment). Using time-integrated quantities of environmental factors to describe
 259 the process of motility, the correct timing of the bloom with an ascending and
 260 descending resting stage may be reproduced. Alternatively, the motility terms

may be replaced by a mechanism that instantaneously transfers all encysting cells into the bottom layer of the model, while excysting cells are analogously transferred to the surface layer.

3.4.4. A four-compartment LCM

Current knowledge on how to adequately subdivide the life cycle of HAB species into distinct stages and how to describe the conditions for stage transitions is sparse. Different species may also require a different number of stages. While the conceptual model of Whipple et al. (2005) identifies 15 life cycle stages for *Phaeocystis*, the life cycle of other species may be captured with sufficient accuracy with fewer stages. For example, Hense and Beckmann (2006) have proposed a prototype schematic that uses two vegetative and two resting stages to describe the life history of cyanobacteria of the order *Nostocales*. The transfer matrix for such a case is expressed as following:

$$\tau_{4 \times 4} = \begin{pmatrix} -\tau_{1,2} & 0 & 0 & \tau_{4,1} \\ \tau_{1,2} & -\tau_{2,3} & 0 & 0 \\ 0 & \tau_{2,3} & -\tau_{3,4} & 0 \\ 0 & 0 & \tau_{3,4} & -\tau_{4,1} \end{pmatrix} \quad (22)$$

This four stage model (Fig. 2D) allows for the discrimination of resting cells that sink (akinetes, the resting stage of cyanobacteria) and rise (germinates). It also treats vegetative cells that take up DIN separate from those that fix dinitrogen gas. Although in the life cycle model by Hense and Beckmann (2006) the transfer between the individual stages are a function of internal quotas (see below), it may be possible to relate the transfer to external factors only (e.g., time integrated quantities).

3.4.5. A twelve-compartment LCM

The consideration of internal quotas are arguably the best way to determine stage transitions. For the four-stage model described in the previous subsection, Hense and Beckmann (2006) have added an energy and a nutrient quota, which leads to a total of 12 compartments. This enables to clearly distinguish the

four stages by their internal quota and in that way to explicitly specify the transfer between the stages (Fig. 3). Thus, if the internal quota of a population approaches a certain (pre-defined) threshold, a transfer into the neighboring stage is induced.

Discrimination between low and high values of two internal quotas allows us to identify each of the four stages with a unique combination of internal states, and to relate the stage succession to changes in internal quotas: For example (Fig. 3), from P_1 (characterized by a high Q_1 and Q_2 -quota) to P_2 (characterized by a high Q_1 and low Q_2 -quota) over P_3 (characterized by a low Q_1 and Q_2 -quota) to P_4 (characterized by a low Q_1 and high Q_2 -quota) and back to P_1 .

A complex multiple compartment LCM has both advantages and disadvantages. Since fully prognostic equations exist for all growing and resting stages, the model can in principle be applied in cases where information about the seed population is missing. It has to be noted, though, that the description of the life cycle (i) is not "mechanistic", (ii) requires a relatively large number of (poorly known) parameters and (iii) may be too expensive to be included in three-dimensional ocean general circulation models.

4. Summary and Conclusions

More than a decade ago, Franks (1997) has presented an overview of the then current harmful algal bloom models, and their representation of biological-physical interaction. At the time, the use of assimilation techniques for forecast purposes seemed most promising. Since then, the importance of the life cycle in HAB dynamics has become clear, and modelers have begun to include life cycle aspects in their models.

Timing, duration, magnitude and distribution patterns of blooms have been found to depend critically on life cycle related processes, in particular excystment and encystment. For example, McGillicuddy et al. (2005) have shown that the inclusion of germination of the benthic resting stages is a prerequisite for

315 obtaining realistic spatial distribution of the toxic dinoflagellate *Alexandrium*
316 *fundyense* in the Gulf of Maine. Hense and Burchard (2009) demonstrated that
317 timing and duration of cyanobacteria blooms in the Baltic Sea are well repre-
318 sented in a full life cycle model, while simpler approaches lead to systematic
319 biases.

320 In order to streamline the various diverse activities of life cycle related HAB
321 modeling, this overview has presented an inventory of the various approaches
322 and listed their main advantages and disadvantages. The focus was on the
323 distinction between growing and resting stages; the methodologies may, however,
324 also be adapted to life cycle transitions concerning for instance colony formation
325 and disruption. A few modeling (e.g., Lancelot et al., 2005) and conceptual
326 studies (e.g., Whipple et al., 2005) have already addressed this topic.

327 The two fundamentally different ways of approaching the time evolution
328 of marine populations (Lagrangian and Eulerian) are both useful for life cycle
329 modeling. While individual-based methods may seem more naturally, fixed
330 grid point (Eulerian) models are equally capable to include life cycle aspects,
331 if subcompartments are introduced, which represent distinct life cycle stages.
332 Preference for one approach over the other should be motivated by the specific
333 goals of the study: Lagrangian methods are well suited for studies related to
334 short term singular events, if explicit time information of individual populations
335 is required and/or in regimes where advection and motility is more important
336 than diffusion. Eulerian methods are advantageous, if a larger domain needs to
337 be uniformly covered and/or if integrated (biogeochemical) quantities are to be
338 determined.

339 In the hierarchy of Eulerian HAB models, it seems that one-compartment
340 LCMs are hardly able to represent the observed bloom dynamics (see Hense
341 and Burchard, 2009), while 1.5-compartment models (Yamamoto et al., 2002;
342 McGillicuddy et al., 2005) do better due to the prescription of boundary condi-
343 tions (abundance and distribution of resting cells). Such models, however, are
344 only semi-prognostic, as the life cycle is not fully closed.

345 It is not clear at the moment, whether two- or four-compartment models

will be a significant step forward (the evaluation is ongoing); but it seems that multi-compartment models (e.g., Hense and Beckmann, 2006) (or a similarly complex Lagrangian model) have the largest potential for capturing the essence of the life cycle dynamics of HABs. Such models, however, have to rely on a large number of (unknown) parameters, which have not yet been determined or confirmed by observations.

Even for key species or key groups we still lack basic understanding of triggering factors for life cycle transition, as well as rates for en- and excystment, metabolism for the distinct stages and biomass losses (e.g., during sex). Thus, many open questions need to be resolved to complete our picture of the life cycle of phytoplankton. Since the importance of the life cycle in regulating HABs is by now obvious, progress in this area is to be expected for the coming years. Modeling activities will certainly play a large part in it.

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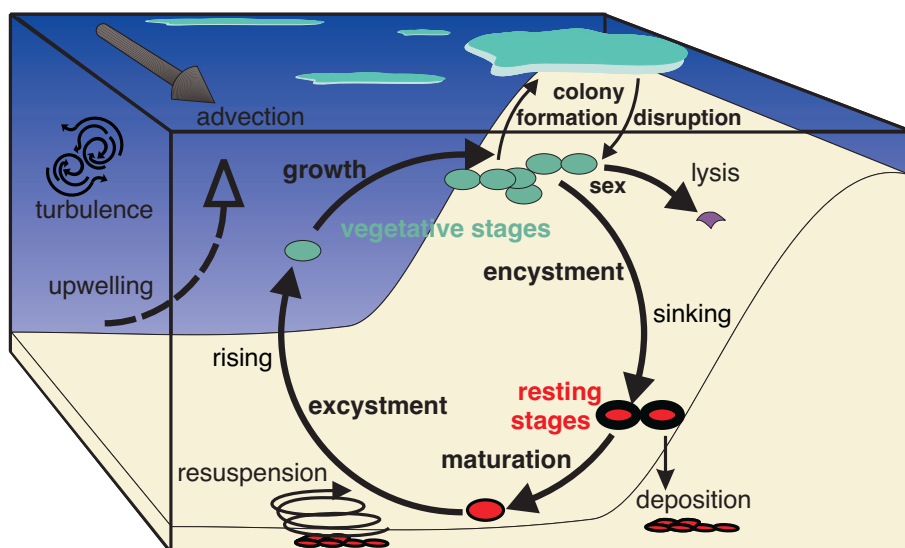


Figure 1: Overview of major life cycle stages of marine phytoplankton, the main biological and contributing physical processes. Beginning with the vegetative phase, cells grow dependent on endogenous and exogenous factors. For some species this may be followed by formation and disruption of colonies. Encystment terminates the vegetative phase and newly formed resting cysts settle down to the sediment. After maturation and possibly resuspension, germination takes place. Subsequent rising of the cells (buoyancy induced, by active upward migration and/or due to upwelling) into the euphotic zone closes the loop. For some (e.g., dinoflagellate, diatom) species sex is involved in life cycle transition (e.g. cyst or colony formation). The spatial distribution of HAB-patches (and fate of the blooms) will depend on the ocean currents and turbulent mixing.

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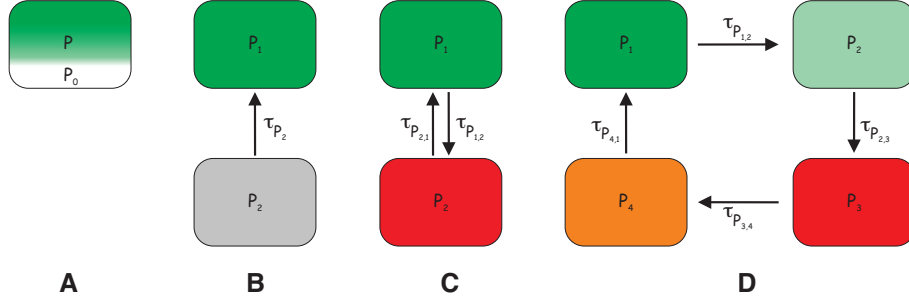


Figure 2: Schematic diagrams of four different Eulerian life cycle modeling approaches (A-D): A) one-compartment LCM which considers a minimum value of phytoplankton (P_0), B) 1.5-compartment LCM which considers a one-way transfer (τ) from a prescribed pool of a seed population (P_2 , obtained, e.g., from observations), C) two-compartment LCM which considers a two-way transfer between the growing (P_1) and the resting stage (P_2), D) four-compartment model which considers two growing and two resting stages which are connected by a unidirectional closed loop.

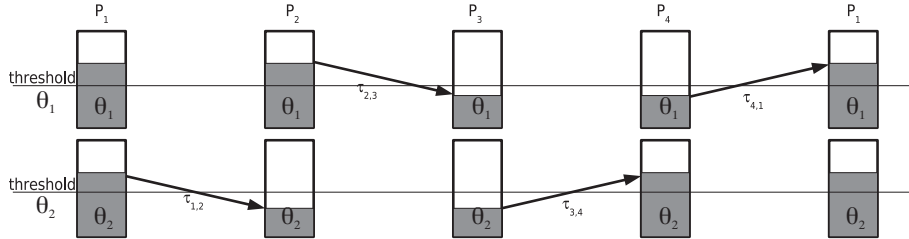


Figure 3: Schematic illustration of the succession of life cycle stages ($P_1 - P_4$) which are characterized by low/high values of two internal quotas (θ_1, θ_2). The arrows indicate the decrease (or increase) of the respective quota below (or above) a certain threshold leading to a transfer into the next stage.